2016 Chehalis ASRP Stream-Associated Amphibian Survey Summary 3rd Progress Report for Post-Feasibility Effort

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EXECUTIVE SUMMARY: The purpose of stream-associated amphibian surveys was to detect and determine the distribution of Dunn's and Van Dyke's salamanders (*Plethodon dunni* and *P. vandykei*) in the Chehalis headwater landscape in and around the footprint of the proposed dam and reservoir. Dunn's and Van Dyke's salamanders are two of non-fish species targeted in the headwater stream-associated guild to examine potential effects of the proposed dam and reservoir on wildlife. This progress report updates the effort begun in 2014 with data from the 2016 season.

We surveyed 37 sites for the 2016 part of this effort, 32 sites were new (not surveyed in either 2014 or 2015). At each site, surveys involved a substrate search of nine 3 m \times 5 m terrestrial plots spaced at 5-10 m intervals with their short axis abutting the wetted stream edge. In 2016, we detected 84 Dunn's salamanders at 18 sites, and 44 Van Dyke's salamanders at nine sites. Four of the Van Dyke's salamander sites were new, the remaining five sites were resurveyed sites where Van Dyke's salamanders had been found in 2014 and 2015; these sites were resurveyed to verify their continued presence. Sixteen of the Dunn's salamander sites were new. The 2016 effort also incidentally recorded 417 individuals representing 10 additional native amphibian species.

Our 2016 effort reinforced patterns we first began to recognize in 2015, which are: 1) Van Dyke's salamander detection increases with elevation; 2) Dunn's salamander detection generally declines with elevation; 3) Van Dyke's salamander is infrequent in the proposed footprint of the dam and reservoir; 4) Dunn's salamander is relatively frequent in the proposed footprint of the dam and reservoir; 5) life history needs of both salamanders will disallow their re-establishment along the "new" riparian margins of the reservoir for either the FRO or the FRFA proposed dam designs; and 6) either dam design will have uncertain isolating effects on Dunn's and Van Dyke's salamanders because disruption of stream-margin connections among populations located in tributaries to the Chehalis mainstem is anticipated as the result of reservoir positioning.

In summary, stream-associated amphibian surveys to date (2014-2016) have surveyed 120 unique sites. Collectively, surveys have made observations of 176 Dunn's salamanders at 41 unique sites, and 86 Van Dyke's salamanders at nine different sites. For Dunn's salamander, this includes 3 of 11 (27%) sites, 24 of 66 (36%) sites, and 13 of 43 (35%) sites, respectively, below, within, and above the footprint of the proposed dam and reservoir. For Van Dyke's salamander, this include 0 of 11 (0%) sites, 1 of 66 (2%) sites, and 8 of 43 (19%) sites, respectively, below, within, and above the same footprint. Breakdown by elevation reveals a sharper difference

between the two species. We found Dunn's salamander at 26 of 70 (37%) sites, 16 of 34 (47%) sites, and 1 of 16 (6%) sites, respectively, at <750 ft (229 m), 750 to 1500 ft (229 to 457 m) and >1500 ft (457 m). In contrast, we found Van Dyke's salamander at 1 of 70 (1%) sites, 4 of 34 (12%) sites, and 4 of 16 (25%) sites over those elevation categories. Across all years, we have also incidentally made 811 observations of at least 10 additional native amphibian species.

INTRODUCTION: This report summarizes the results of the Chehalis ASRP stream-associated amphibian surveys to date (through 2016 sampling) in headwaters of the Chehalis mainstem that encompasses the vicinity of the proposed footprint of the dam and its reservoir and the adjacent headwater areas. These surveys focus on the terrestrial stream-associated amphibians because Dunn's salamander (Plethodon dunni) and Van Dyke's salamander (P. vandykei), two of the eight ASRP non-fish aquatic-habitat associated target species, are regarded as stream-associated but occur in the terrestrial riparian habitat immediately adjacent streams. We first initiated stream-associated amphibian surveys on 24 February 2014. This progress report augments information from surveys completed in 2014 and 2015. In 2016, surveys began on 6 March and were completed on 5 May 2016. Several years are required to complete an adequate assessment because the time window in which these species (particularly Van Dyke's salamander) are surface active (that is, effectively detectable) in any one year is short (about three months), which limits the number of sites that can be effectively surveyed annually.

SITE SELECTION: We chose sites from a 187-site pool systematically placed along the stream network within the coniferous forest landscape to provide an array dispersed across the footprint of the proposed dam and its reservoir and the surrounding area. When we discuss the footprint of the proposed dam and reservoir, footprint means the reservoir at full pool. Besides the fact that this footprint includes the proposed dam and reservoir area, this footprint was selected based on the coniferous forest landscape being the only Chehalis Basin habitat where Van Dyke's salamander might be expected. As in previous years, we selected sites in this pool at a minimum distance 400 m from one another to ensure independence among sites; the annual movement scale of terrestrial salamanders like Dunn's and Van Dyke's salamanders is <20 m (Ovaska 1988, Staub et al. 1995).

In 2014, we developed an original pool of 128 sites from which to select sites, but increased that pool by 15 new sites later in 2014, by 43 new sites in 2015, and by one new site in 2016. One reason for augmenting the site pool was our inability to access selected sites because of washed out roads or because selected sites lacked habitat for sampling (e.g., a bedrock cliff face) or had too little habitat to lay out a sampling array (see Sampling section). This reason led us to reject five sites in 2014, nine sites in 2015, and 25 sites in 2016. The other reason for adding to the site pool was to enable selecting enough sites outside the proposed dam and reservoir footprint to understand potentially different patterns of distribution within versus outside of that footprint. We considered the latter critical to characterizing the consequence of potential changes in habitat resulting from the location of a proposed dam/reservoir footprint.

In 2014, we selected sites to be sampled so that about 60% of the sites were from within the proposed dam/reservoir footprint; the remaining about 40% of the sites were selected from above and below the dam/reservoir footprint in a ratio of 9:1 above versus below the reservoir. In 2015 and 2016, we intentionally created the target of 14 sites to be sampled outside of the footprint that were selected in a ratio 11:3 above versus below it. We designed this selection pattern to capture potential differences in species distribution that might occur as a consequence of footprint location when considering the large area of coniferous forest habitat available upstream, which is similar to existing habitat within the footprint. In contrast, coniferous forest habitat was very limited downstream of the footprint. In 2015 and 2016, we intentionally shifted the ratio of sites sampled upstream versus downstream of the proposed dam and reservoir to ensure enough sites would exist downstream to enable effective comparison to other areas. Figure 1 shows both sites available for sampling and sites sampled in 2016; sites sampled within the proposed dam/reservoir footprint are gold, those below the footprint are pink and those above the footprint are in blue. Sites in the pool but that we did not select are in white .

We had a target minimum number of sites to be sampled each year based on our effective sampling window and field crew size. Our minimum target number of sites was 30 in 2016 and 2015, and 51 in 2014. In each of 2015 and 2016, our target distribution of sites was 16 within, 11 above, and three below the proposed dam footprint; in 2014, our target distribution was 31 within, 18 above, and 2 below the proposed dam footprint. Further, in 2016 we resampled the five sites where Van Dyke's salamanders were found in 2014 and 2015 to assess whether the species was still present at those sites; and in 2015, we resampled the two sites where Van Dyke's salamanders were found in 2014 for the same reason.

SAMPLING: We conducted all surveys with a field crew of at least three with sampling done on four days each week. We surveyed by laying out nine 3 m wide × 5 m long plots at each site, each of which abutted the wetted edge of the stream along their short axis. Sampling was done by raking through the litter (leaves, conifer needles, and small wood debris), rock and soil substrate with a potato rake, overturning movable surface objects, and taking apart woody debris sufficiently decayed to be dismantled.

RESULTS: In 2016, we exceeded our planned 30-site survey target with 37 sites surveyed, including resurvey of the five unique sites where Van Dyke's salamander was in 2014 and 2015 (**Table 1**). Four of the 37 sites were located below the footprint of the proposed dam and reservoir, 17 sites were located within the footprint, and 16 sites were located upstream of the footprint (**Figure 1**). Resurveyed Van Dyke's salamander sites included one within the footprint of the proposed dam and reservoir and four upstream of the footprint.

In 2014, we had exceeded our planned 51-site target with 58 sites surveyed (**Table 1**). Four of the 58 sites were located below the inundation footprint of the proposed dam and reservoir, 34 sites were located within the footprint, and 20 sites were located upstream of the footprint (**Appendix Figure 1**).

Table 1. Target site numbers, numbers of sites actually sampled and their distribution relative to the proposed dam and reservoir footprint and Van Dyke's sites resurveyed by year. Unique represents the total number of different sites summed across all years. Sampled represents the total number of sites sampled including sites that were repeat-sampled.

Year	Site N	umbers	<u>Locatio</u>	n Relative to Fo	Van Dyke's Sites			
	Target	Sampled	Below	Within	Resampled			
2014	51	58	4	34	20	-		
2015	30	32	3	17	12	2		
2016	30	37	4	17	16	5		
Unique	111	120	11	66	43	-		
Sampled	-	127	11	68	48	7		

In 2015, we had exceeded our planned 30-site target with 32 sites surveyed, including the two sites sampled in 2014 for Van Dyke's salamander (**Table 1**). Three of the 32 sites were located below the inundation footprint of the proposed dam and reservoir, 17 sites were located within the footprint, and 12 sites were located upstream of the footprint (**Appendix Figure 1**). The resurveyed Van Dyke's salamander sites included one within the footprint of the proposed dam and reservoir and one upstream of the footprint.

Over the three years to date (2014-2016), we exceeded our planned 111-site target with 120 unique sites surveyed (**Table 1**). Eleven unique sites were located below the inundation footprint of the proposed dam and reservoir, 66 unique sites were located within the footprint, and 43 unique sites were located upstream of the footprint (**Figure 1** and **Appendix Figures 1-2**).

In 2016, we recorded observations of 545 individuals of 12 species of amphibians at the 37 sites sampled (**Table 2**, **Figures 2a**, **2b and 2c**). We found at least one species of amphibian at all 37 sites. The four species of terrestrial amphibians (all salamanders) recorded represent 69.5% of observations; the 8 non-terrestrial amphibian species we recorded represented 30.5% of observations.

In 2014, we recorded observations of 337 individuals of up to 12 species of amphibians at the 58 sites sampled (**Appendix Table 1, Appendix Figures 2a, 2b and 2c**). We found at least one species of amphibian at 51 of the 58 sites. The four species of terrestrial amphibians recorded represent 72.7% of observations; the 8 non-terrestrial species we recorded represented 27.3% of observations.

Figure 1. Distribution for Sites for Stream-Associated Amphibians Surveys. Proposed survey sites include both sites not selected and sites rejected for the reasons specified in the text.

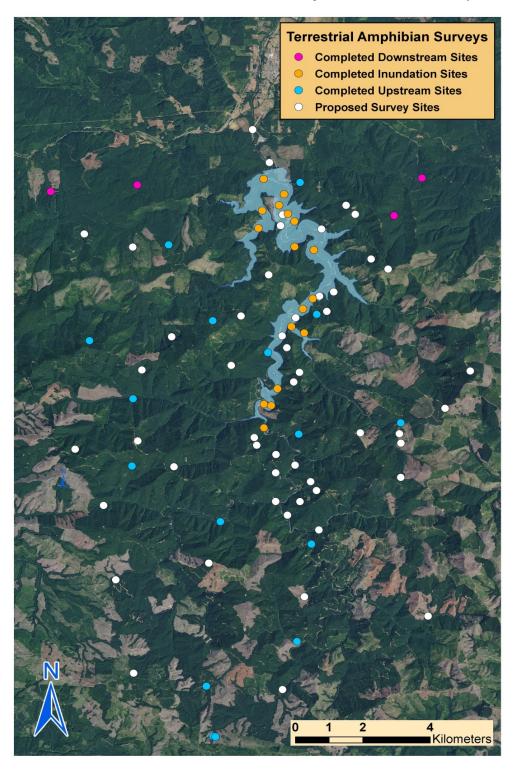


Table 2. Amphibian species and numbers of observations during terrestrial amphibians surveys, March-May 2016. Subtotals or totals for sites may be less than summed site sums for species across habitat categories because one or more species may have occurred at the same site. The overall number of sites includes two sites with incidental observations.

Species			Numbers of Sites and Individuals (Ind) observed							
			Below		In		Above		Totala	
Standard English Name	Scientific Name	footprint		footprint		footprint		Totals		
		Sites	Ind	Sites	Ind	Sites	Ind	Sites	Ind	
Terrestrial Amphibians										
Dunn's salamander	Plethodon dunni	1	1	10	44	7	39	18	84	
Ensatina	Ensatina eschscholtzii	3	5	4	10	3	3	10	18	
Van Dyke's salamander	Plethodon vandykei	0	0	1	2	8	42	9	44	
Western red-backed salamander	Plethodon vehiculum	4	13	15	95	14	125	33	233	
	Subtotals	4	19	15	151	16	209	35	379	
Stillwater-breeding Amphibians										
Pacific treefrog	Pseudacris regilla	0	0	5	7	1	1	6	8	
Northern red-legged frog	Rana aurora	1	1	1	1	3	3	5	5	
Northwestern salamander	Ambystoma gracile	2	4	1	1	1	1	4	6	
Roughskin newt	Taricha granulosa	1	1	2	2	3	5	6	8	
Western toad	Anaxyrus boreas	0	0	8	21	3	3	11	24	
Subtotals			6	10	32	9	13	22	51	
Stream-breeding Amphibians										
Giant salamanders	Dicamptodon tenebrosus	1	1	0	0	5	13	6	14	
Coastal tailed frog	Ascaphus truei	1	1	1	1	7	16	9	18	
Columbia torrent salamander	Rhyacotriton kezeri	1	3	6	30	11	50	18	83	
	Subtotals	3	5	7	31	13	79	23	115	
Overall Totals			30	17	214	16	301	37	545	

Figure 2a. Distribution of terrestrial amphibians, including stream-associated taxa, encountered in 2016. Numbers of specific taxa are summarized adjacent to either individual points or boxes encompassing groups of individual points.

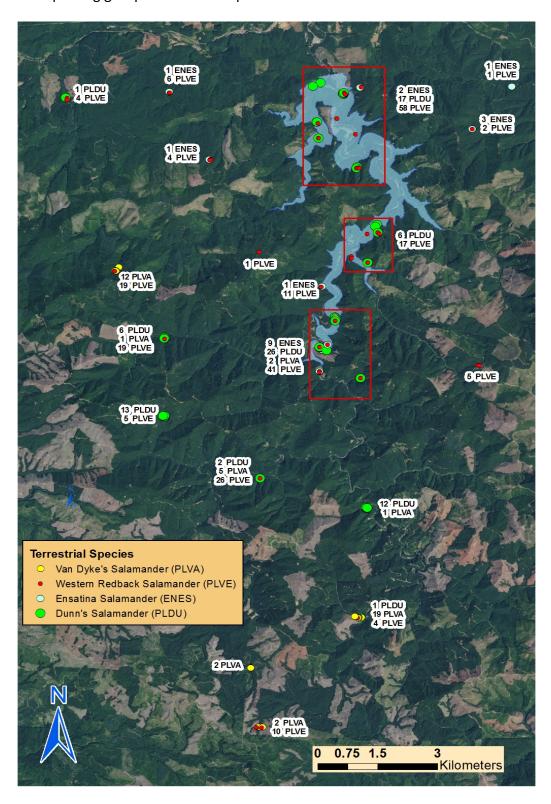


Figure 2b. Distribution of stillwater-breeding amphibians incidentally recorded during stream-associated amphibian surveys in 2016. Numbers of specific taxa are summarized next to individual points or boxes encompassing groups of individual points.

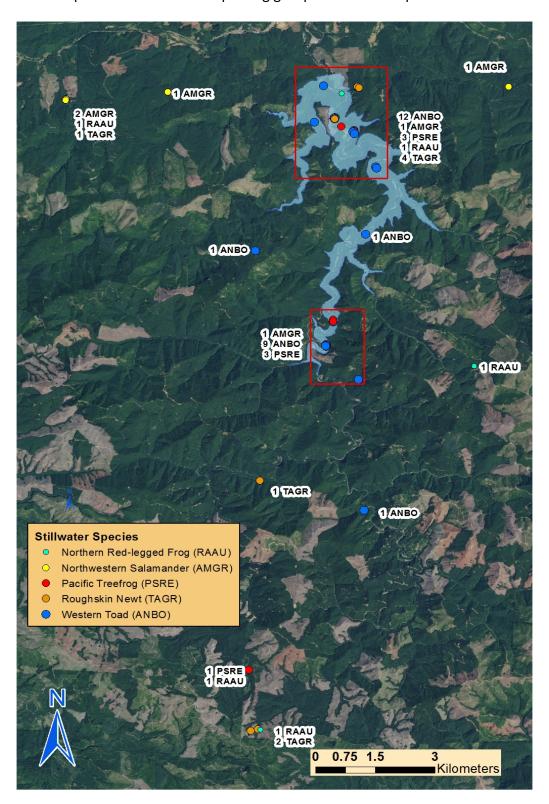
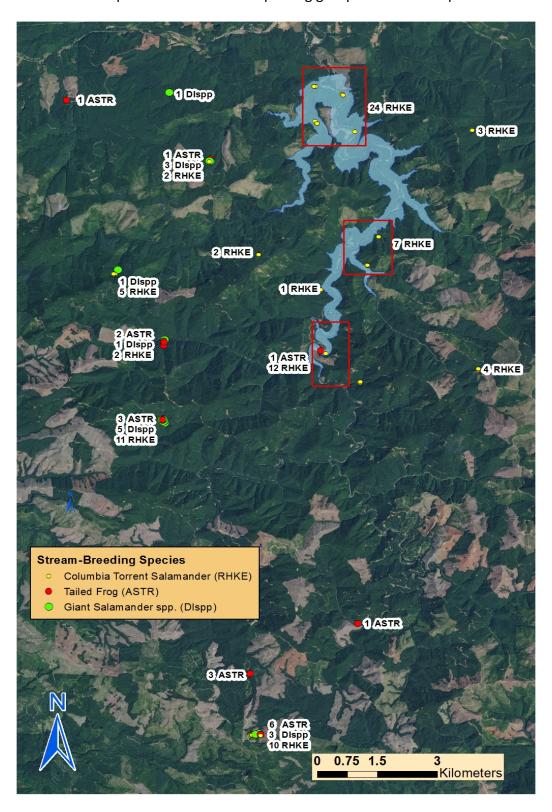


Figure 2c. Distribution of stream-breeding amphibians incidentally recorded during stream-associated amphibian surveys in 2016. Numbers of specific taxa are summarized adjacent to either individual points or boxes encompassing groups of individual points.



In 2015, we recorded observations of 354 individuals of at least 10 species of amphibians at the 32 sites sampled (**Appendix Table 2, Appendix Figures 3a, 3b and 3c**). We found at least one species of amphibian at all 32 sites. The four species of terrestrial amphibians (all salamanders) recorded represent 81.9% of observations; the 8 non-terrestrial species we recorded represented 18.1% of observations.

Over the three years to date (2014-2016), we recorded observations of 1,236 individuals representing at least 12 species of amphibians at the 120 unique sites surveyed. We found at least one species of amphibian at every site. The four species of terrestrial amphibians (all salamanders) recorded represent 74.7% of all observations; the 8 non-terrestrial species we recorded represented 25.3% of observations.

In 2016, western red-backed salamanders (*Plethodon vehiculum*) were the most frequently encountered terrestrial salamander species, representing 61.5% of all observations and recorded at 89.2% of sites. The second most frequently encountered was Dunn's salamander, representing 22.2% of observations and recorded at 48.6% of sites. Van Dyke's salamander and Ensatina (*Ensatina eschscholztii*) were the two least frequently recorded species, being found, respectively, at 24.3% and 27.0% of sites and representing 11.6% and 4.7% of observations. Van Dyke's and western red-backed salamanders were the only two terrestrial amphibian species recorded at more sites and had more animals above the footprint of the proposed dam and reservoir than anywhere else. The four terrestrial amphibians also differed in the mean number of individuals recorded per site sampled. For sites at which they were found, western red-back salamanders had the highest mean number of individuals per site (7.1 individuals/site), followed by Van Dyke's salamanders (4.9 individuals/site), Dunn's salamanders (4.7 individuals/site), and Ensatina (1.8 individuals/site).

In 2014, western red-backed salamanders had also been the most frequently encountered terrestrial salamander species, representing 78.0% of all observations and recorded at 72.5% of sites (Appendix Table 1). The second most frequently encountered was Ensatina, representing 24.1% of sites, followed by Dunn's salamander, representing 9.8% of observations and recorded at 17.2% of sites. Van Dyke's salamander was the least frequently recorded species, recorded at 3.4% of sites and representing 4.9% of observations. The four terrestrial amphibians also differed in the mean number of individuals recorded per site sampled. For sites at which they were found, Van Dyke's salamanders had the highest mean number of individuals per site (6.0 individuals/site), followed by Western red-back salamanders (5.2 individuals/site), Dunn's salamanders (2.4 individuals/site), and Ensatina (1.3 individuals/site).

In 2015, western red-backed salamanders had also been the most frequently encountered terrestrial salamander species, representing 63.1% of all observations and recorded at 81.3% of sites (Appendix Table 2). The second most frequently encountered was Dunn's salamander, representing 23.4% of observations and 46.9% of sites. Ensatina and Van Dyke's salamander were the two least frequently recorded species, being found, respectively, at 18.8% and 15.6% of sites and representing 3.1% and 10.3% of observations. For sites at which they were found, the four terrestrial amphibians also differed in the mean number of individuals recorded per

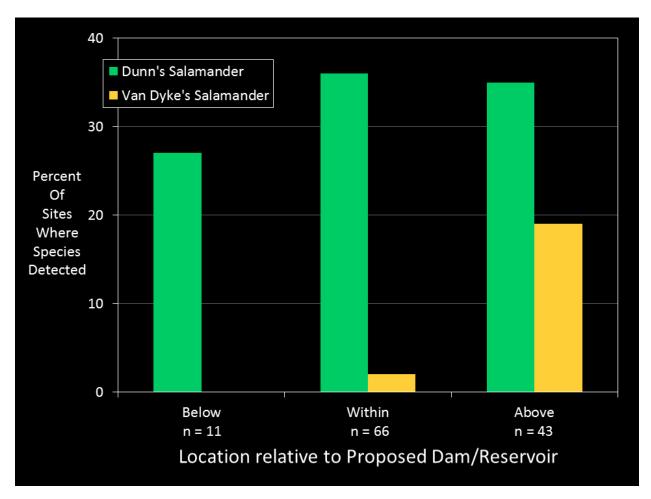
site sampled. Western red-back salamanders had the highest mean number of individuals per site (7.0 individuals/site), followed by Van Dyke's salamanders (6.0 individuals/site), Dunn's salamanders (4.5 individuals/site), and Ensatina (1.5 individuals/site).

Overall (2014-2016), western red-backed salamanders (*Plethodon vehiculum*) were the most frequently encountered of terrestrial salamanders observed, representing 49.1% of all observations and recorded at 77.5% of sites. The second most frequently encountered was Dunn's salamander, representing 14.2% of observations and recorded at 34.2% of sites. Van Dyke's salamander and Ensatina were the two least frequently recorded species, being found, respectively, at 7.5% and 25.0% of sites, and representing, respectively, 7.0% and 3.6% of observations. Van Dyke's had more animals above the footprint of the dam than anywhere else. The four terrestrial amphibians also differed in the mean number of individuals recorded per site sampled. Western red-back salamanders had the highest mean (6.5 individuals/site), followed by Van Dyke's salamanders (5.6 individuals/site), Dunn's salamanders (3.9 individuals/site), and Ensatina (1.5 individuals/site). Van Dyke's salamanders were found at all five sites where they were recorded in 2014 and 2015 and four new sites above the proposed dam footprint.

Dunn's and Van Dyke's salamanders are also differentially distributed relative to the location of the proposed dam/reservoir footprint and elevation. Based on location relative to the footprint, we found no Van Dyke's salamanders below the footprint and at only one site within the footprint (2% of 66 unique sites sampled within the footprint); all remaining Van Dyke's salamander locations (n = 8) were above the footprint and represented 19% of the 43 unique sites sampled above the footprint (Figure 3). In contrast, Dunn's salamander was found with greater frequency below, within and above the footprint with the percent of sites where the species was detected ranging from 27% of 11 unique sites sampled below the footprint to 36% of 66 unique sites within the footprint to 30% of 43 unique sites above the footprint. However, location partitioning relative to the proposed dam/reservoir footprint conceals a prominent pattern relative to elevation. Partitioning into 750-ft elevation blocks reveals that Dunn's salamander is roughly equally distributed in distribution, measured as the proportion of sites where detected, in the two lowest elevations blocks, but drops off markedly in the highest elevation block (Figure 4). In contrast, Van Dyke's salamander was not detected in the lowest elevation, was rare in the mid-elevation block (found at one site), and the remaining sites were it was detected were in the highest elevation block. However, in contrast to Dunn's salamander, the frequency with which Van Dyke's salamander was detected across the landscape was consistently lower.

Discussion: With few exceptions, patterns we observed in 2016 are similar to those observed during surveys in years 2014 and 2015. As expected, since the surveys were designed to detect terrestrial salamanders, most amphibians detected were terrestrial salamanders. In particular,

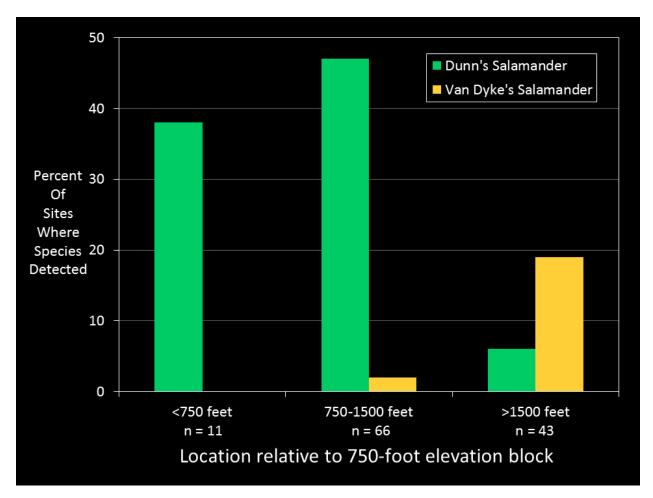
Figure 3. Distribution of Dunn's and Van Dyke's salamanders relative to the footprint of the proposed dam and reservoir. Number of unique sites sampled in each location category are indicated below category labels.



terrestrial salamander observations (n = 914 across all years) represented 69.5% to 81.9% (\bar{x} = 74.7%) of amphibians found across each of the three years. However, a substantial number (n = 322) of amphibians representing 11 additional species were recorded incidentally in these surveys. This richness underscores the high amphibian species richness of the Chehalis Basin, which is the highest in Washington State and equaled only by a few areas in the South Cascades (Dvornich et al. 1997). This richness occurs in a landscape that is entirely timber-managed.

Perhaps the most striking difference in 2016 data is the higher numbers of amphibians per sites encountered when contrasted to 2014 and 2015. In particular, we found 14.7 amphibians per sample site in 2016, this contrasts to 5.8 amphibians/site in 2014 and 11.0 amphibians per site in 2015. This pattern likely reflects greater terrestrial salamander surface activity under wetter conditions; the October 2015-March 2016 interval was regionally the wettest in the historical record.

Figure 4. Distribution of Dunn's and Van Dyke's salamanders relative to elevation. Number of unique sites sampled in each elevation category are indicated below category labels.



Western red-backed salamander was the most frequently recorded terrestrial amphibian; this agrees with previous work on Western red-backed salamanders, which require relatively mesic terrestrial habitats, are typically the most frequently recorded terrestrial salamander in the generally more mesic Willapa Hills (M. Hayes, unpublished data) as well as generally in Coast Ranges habitats in Washington (Raphael et al. 2002).

Ensatina, a relatively drier habitat-adapted terrestrial salamander species, was much less frequent than the Western red-backed salamander in this mesic Coast Range habitat, a pattern also recorded elsewhere (Raphael et al. 2002). Ensatina tends to be more frequent in more interior, forested Oregon and Washington (Bury et al. 1991). Further, our surveys were riparian-focused to enable detecting Dunn's and Van Dyke's salamanders, they would be expected to be less frequent in the riparian margin than in the drier adjacent uplands. We recorded Ensatina at a slightly greater proportion of sites in 2014 (24.1%) & 2016 (27.0%) in contrast to 2015 (18.8%). Opportunistic draw within in our stratified random selection of sites may be the basis as analysis among those proportions reveals them to not differ significantly.

Dunn's salamander was more frequently recorded than Ensatina, which undoubtedly simply reflects the riparian-focused nature of our surveys. Dunn's salamander, a terrestrial salamander with greater moisture requirements than Ensatina, is a more stream-associated terrestrial species and the terrestrial amphibian surveys were stream margin-focused. We did not survey the less mesic uplands away from the stream, where more Ensatina might be expected. Dunn's salamander was relatively frequently recorded within the footprint of the proposed dam and reservoir (Figure 3), but appears to be less frequent at higher elevations (Figure 4).

Van Dyke's salamander was infrequently recorded. Van Dyke's salamander, also strongly stream-associated species, is the least frequently recorded terrestrial salamander in several Coast Range habitats and other studies in western Washington (Wilkins and Peterson 2000, Raphael et al. 2002, McIntyre 2003, Wilk et al. 2014). Only two historical records exist for Van Dyke's salamander from the upper Chehalis system (WDFW WSDM database, accessed 12 February 2014). Both records originated from the studies that Weyerhaeuser conducted when they developed their Landscape Conservation Plan across the Willapa Hills landscape in the 1990s and both come from what is best described as mid-elevation in the Chehalis headwaters (around 1300 feet [400 m]). We found Van Dyke's salamander at four new sites in 2016, all of which were above the proposed footprint of the dam. Though Van Dyke's salamander has been found at few sites overall (n = 9), the distribution of sites at which it was found indicates that the species is more frequent above (n = 8) than within (n = 1) the proposed dam/reservoir footprint. However, the pattern is consistent with this species being a cool-adapted stenotherm, since the old Forest Service Survey and Manage Species criteria for Van Dyke's salamander recommend that surveys be conducted at air temperatures ≤15°C [59°F] (Jones 1999), and temperatures that satisfy its presumed optimal thermal regime are more frequent at the higher elevations.

The differential pattern between Dunn's and Van Dyke's salamanders indicates that in the event of construction of either an FRO or FRFA dam option, Dunn's salamander would lose significant habitat with the footprint of either option. In contrast, Van Dyke's salamander would proportionally lose much less suitable habitat than Dunn's salamander. Both the FRO and FRFA options are anticipated to eliminate riparian habitat with the structure that could currently support either Dunn's or Van Dyke's salamanders because vegetation removal in the reservoir footprint is expected under either option. Since the habitat for both these salamanders is exclusively riparian, the reservoir footprints for either dam option are likely to also isolate Dunn's and Van Dyke's salamander populations in tributaries to the Chehalis mainstem where habitat for both species is anticipated to remain. What that isolation will do to locate gene flow in populations of both these species is unknown, and a condition of high uncertainty.

We should also note that in the anticipated climate change trajectory, we would expect Van Dyke's salamander to retreat upwards in elevation, where Dunn's salamander would be expected to expand upwards in elevation. However, at a certain point, Dunn's expansion upwards in elevation may be accompanied by a retreat along its lower elevation limit. Both

these patterns are excellent justification for including acquisitions in the headwater portions of this system to help attenuate the climate change trajectory as part of restoration options.

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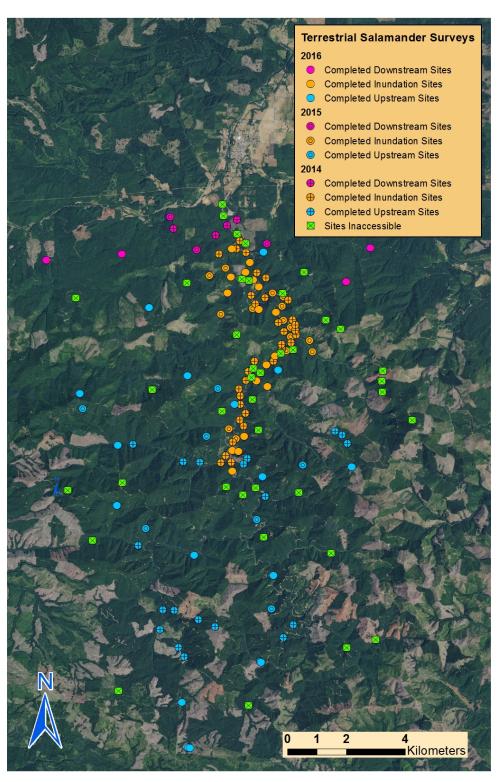
Appendix Table 1. Amphibian species and numbers of observations during terrestrial amphibians surveys, February-July 2014. Subtotals or totals for sites may be less than summed site sums for species across habitat categories because one or more species may have occurred at the same site.

Species			Numbers of Sites and Individuals (Ind) observed							
	Scientific Name	Below		In		Above		Totals		
Standard English Name		footprint		footprint		footprint				
		Sites	Ind	Sites	Ind	Sites	Ind	Sites	Ind	
Terrestrial Amphibians										
Dunn's salamander	Plethodon dunni	0	0	7	16	3	8	10	24	
Ensatina	Ensatina eschscholtzii	1	1	5	6	8	11	14	18	
Van Dyke's salamander	Plethodon vandykei	0	0	1	5	1	7	2	12	
Western red-backed salamander	Plethodon vehiculum	3	7	18	93	16	91	37	191	
Subtotals			8	20	120	17	117	40	245	
Stillwater-breeding Amphibians										
Pacific treefrog	Pseudacris regilla	0	0	6	11	0	0	6	11	
Northern red-legged frog	Rana aurora	1	1	1	1	1	1	3	3	
Roughskin newt	Taricha granulosa	0	0	2	2	1	2	3	4	
Western toad	Anaxyrus boreas	1	1	6	10	2	2	9	13	
Subtotals			2	20	24	4	5	27	31	
Stream-breeding Amphibians										
Giant salamanders	Dicamptodon sp.	0	0	0	0	4	4	4	4	
Coastal tailed frog	Ascaphus truei	0	0	5	6	4	6	9	12	
Columbia torrent salamander	Rhyacotriton kezeri	0	0	6	34	7	11	13	45	
	Subtotals	0	0	11	40	9	21	20	61	
Overall Totals			10	23	184	22	143	48	337	

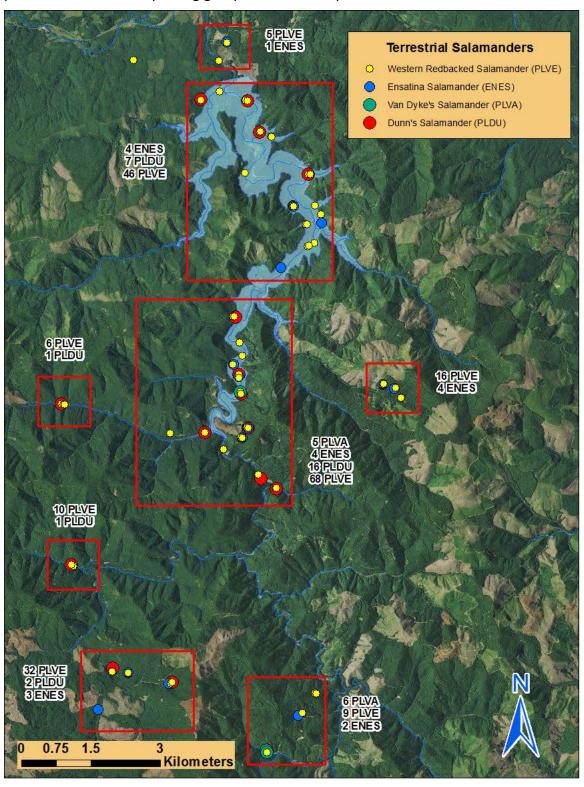
Appendix Table 2. Amphibian species and numbers of observations during terrestrial amphibians surveys, March-April 2015. Subtotals or totals for sites may be less than summed site sums for species across habitat categories because one or more species may have occurred at the same site.

Species			Numbers of Sites and Individuals (Ind) observed							
Standard English Name	Scientific Name	Below footprint		In footprint		Above footprint		Totals		
_		Sites	Ind	Sites	Ind	Sites	Ind	Sites	Ind	
Terrestrial Amphibians										
Dunn's salamander	Plethodon dunni	2	3	9	36	4	29	15	68	
Ensatina	Ensatina eschscholtzii	1	2	3	4	2	3	6	9	
Van Dyke's salamander	Plethodon vandykei	0	0	1	3	4	27	5	30	
Western red-backed salamander	Plethodon vehiculum	3	5	13	73	10	105	26	183	
Subtotals			10	15	116	10	164	27	290	
Stillwater-breeding Amphibi	ans			•						
Pacific treefrog	Pseudacris regilla	1	3	2	5	1	2	4	10	
Northern red-legged frog	Rana aurora	0	0	0	0	0	0	0	0	
Roughskin newt	Taricha granulosa	1	6	1	1	0	0	2	7	
Western toad	Anaxyrus boreas	0	0	7	7	1	1	8	8	
Subtotals			9	8	13	2	3	12	25	
Stream-breeding Amphibians										
Coastal giant salamanders	Dicamptodon tenebrosus	0	0	0	0	1	2	1	2	
Coastal tailed frog	Ascaphus truei	0	0	1	2	4	5	5	7	
Columbia torrent salamander	Rhyacotriton kezeri	1	2	5	19	4	9	10	30	
	Subtotals	1	2	5	21	5	16	11	39	
Overall Totals			21	18	150	10	183	32	354	

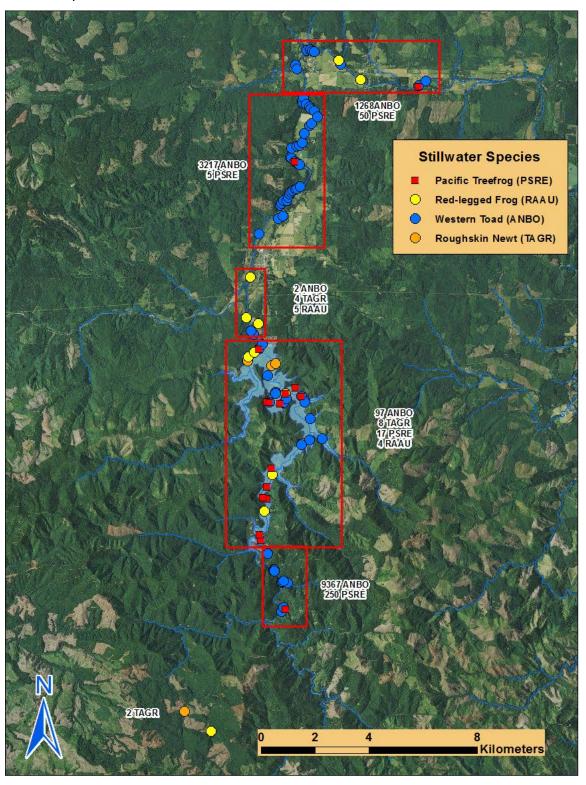
Appendix Figure 1. Location of sampled stream-associated amphibian survey sites by year. Color coding indicates whether sites were downstream of the proposed dam and reservoir footprint (downstream), within the proposed dam and reservoir footprint (inundation) or above the proposed dam and reservoir footprint (upstream). Rejected sites shown in green.



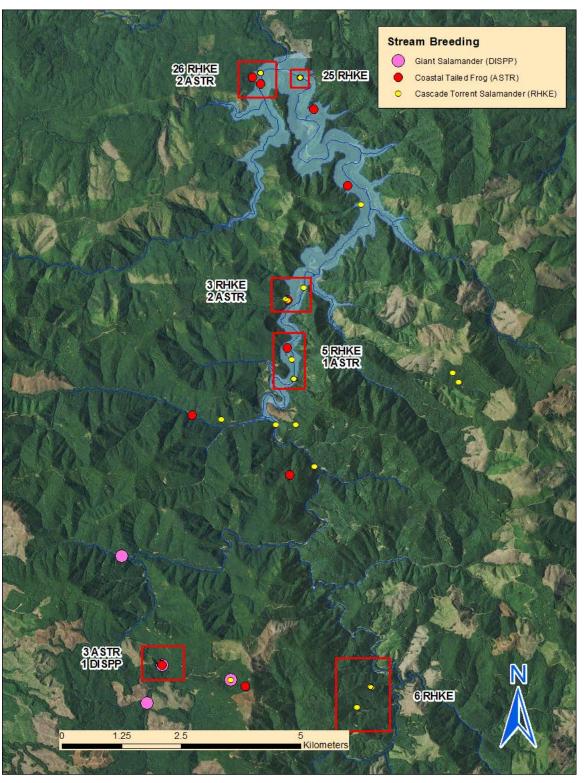
Appendix Figure 2a. Distribution of terrestrial amphibians, including stream-associated taxa, encountered in 2014. Numbers of specific taxa are summarized adjacent to either individual points or boxes encompassing groups of individual points.



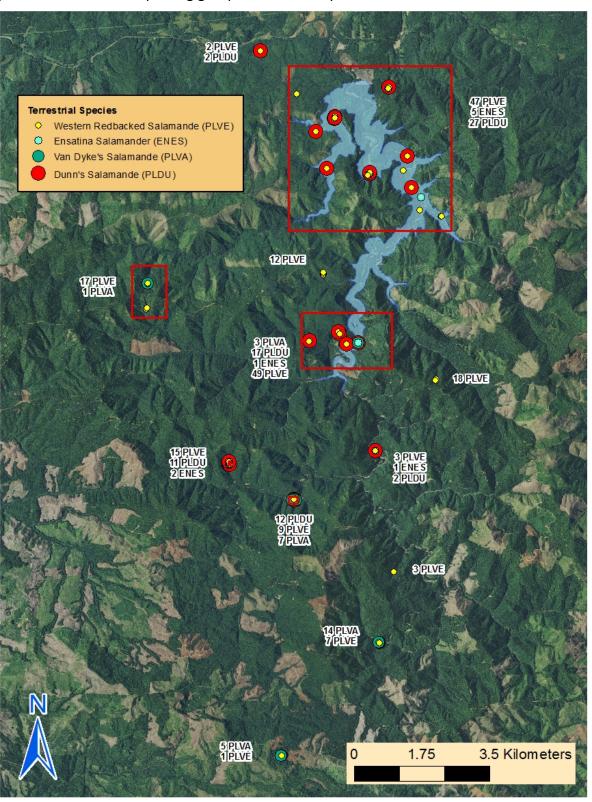
Appendix Figure 2b. Distribution of stillwater-breeding amphibians incidentally recorded during stream-associated amphibian surveys in 2014. Numbers of specific taxa are summarized next to individual points.



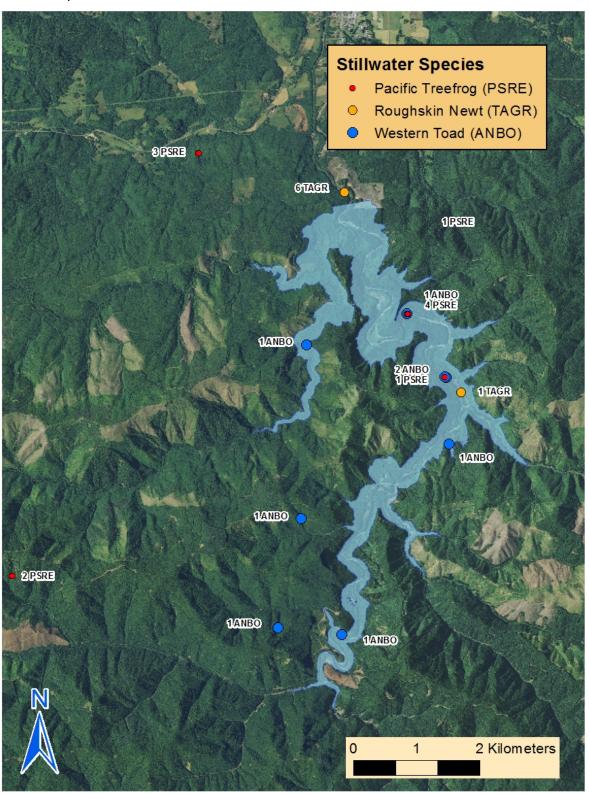
Appendix Figure 2c. Distribution of stream-breeding amphibians incidentally recorded during stream-associated amphibian surveys in 2014. Numbers of specific taxa are summarized adjacent to either individual points or boxes encompassing groups of individual points.



Appendix Figure 3a. Distribution of terrestrial amphibians, including stream-associated taxa, encountered in 2015. Numbers of specific taxa are summarized adjacent to either individual points or boxes encompassing groups of individual points.



Appendix Figure 3b. Distribution of stillwater-breeding amphibians incidentally recorded during stream-associated amphibian surveys in 2015. Numbers of specific taxa are summarized next to individual points.



Appendix Figure 3c. Distribution of stream-breeding amphibians incidentally recorded during stream-associated amphibian surveys in 2015. Numbers of specific taxa are summarized adjacent to either individual points or boxes encompassing groups of individual points.

